

## 2.23

# Government and Agency Response to Soil Erosion Risk in Europe

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### 2.23.1 INTRODUCTION

'The dust is gold that bears the harvest;  
Save the soil that grows our bread;  
Let not wind and rain remove it;  
Guard with care for years ahead.'

SG Brade-Birks (1944)

1 Soil conservation in Europe has not generally received sufficient attention, until recently. However, a picture  
2 has gradually emerged, largely developed by individual research teams from various academic institutions. In  
3 Northern Europe, the Katholieke Universiteit Leuven (Belgium) has made a particular contribution. Erosion  
4 surveys were often conducted by Quaternary geologists and geomorphologists, who increasingly viewed  
5 evidence of agriculturally induced erosion in their landscapes (Fullen, 2000, 2003).

6 Faced with evidence of increased frequency and severity of soil erosion, a considered approach to soil  
7 conservation is essential. This must take into account both the common attributes and diversity of European  
8 agronomic, environmental and social conditions. The soil conservation policies of 10 European countries  
9 (Belgium, Denmark, France, Germany, Hungary, Iceland, Italy, The Netherlands, Spain and the UK) are  
10 reviewed. These countries were selected as representative of European regions (i.e. north, east, west and  
11 Mediterranean Europe, plus Scandinavia). The soil conservation policies are compared with several Australian  
12 and US strategies and several policy instruments recommended as appropriate approaches to soil conservation  
13 throughout Europe.

## 16 2.23.2 REVIEW OF EUROPEAN NATIONAL POLICIES

### 18 2.23.2.1 Belgium

19 Individual regions of Belgium are responsible for their own environmental issues; thus soil conservation  
20 policies differ between the Flanders and Wallonia regions. For instance, soil degradation has been  
21 recognised as an environmental problem for more than a decade in Flanders, resulting in the conception  
22 and implementation of the Flemish Environmental Action Plan (1997–2001). In contrast, policy develop-  
23 ments have not yet been observed in Wallonia. Although several agri-environmental measures (e.g. cover  
24 crops and grass buffer strips along rivers and field boundaries) are implemented by many farmers  
25 experiencing erosion problems, none were originally meant to control erosion as part of regional policy  
26 (Biielders *et al.*, 2003).

27 The latest Flemish Environmental Action Plan (2003–07) recognizes erosion as a threat to long-term soil  
28 sustainability and addresses soil protection through erosion control. In December 2001, the Flemish  
29 Government issued a decree concerning ‘the subsidy of small-scale erosion control measures to be taken  
30 by local authorities’, often called the ‘Soil Erosion Decree’. This Decree regulates subsidies to municipalities  
31 on two levels. First, municipalities dealing with erosion problems in the hilly regions of Flanders are  
32 stimulated to draw up a municipal soil erosion action plan, indicating their proposed measures. The Flemish  
33 Government provides municipalities with €12.5 ha<sup>-1</sup> of plan area. Once the Land Division of the Flemish  
34 Administration (AMINAL, Land) approves a plan, it grants a 75 % subsidy for the implementation of approved  
35 measures. The measures are selected from a ‘Code of Good Practice – Erosion Control’, currently being  
36 prepared by the Flemish Administration, and will include mostly small-scale technical control methods, such  
37 as the construction of small dams, sedimentation ponds, grass buffer strips and grassed waterways. Soil  
38 conservation measures carried out by farmers, such as conservation tillage or cover crops, are not subsidised  
39 via the ‘Erosion Decree’, but can be recommended in a municipal soil erosion action plan. However, for other  
40 environmental reasons, farmers can be subsidised annually for sowing cover crops during autumn (€50 ha<sup>-1</sup>)  
41 and for sowing grass strips with a width of 5–10 m along rivers and sunken lanes (€0.13 m<sup>-2</sup>). These practices  
42 are now also promoted within the framework of the soil erosion control policy. Other subsidies for farmers,  
43 relevant for soil erosion prevention, include subsidies for afforestation (between €850 ha<sup>-1</sup> for poplar and  
44 €3700 ha<sup>-1</sup> for oak) and set-aside (€298–424 ha<sup>-1</sup>, depending on agricultural region and crop type). Further  
45 grants for agricultural practices which decrease soil erosion are expected to be available in 2004, through the  
46 ‘Flemish Plan for Rural Development’, within the framework of European Directives. These measures will  
47

1 include soil conservation actions at the field scale and were approved by the European Commission (EC) in  
2 August 2003.

3 Establishment of a municipal soil erosion action plan occurs in close cooperation with farmers and this  
4 participatory approach has been applied recently in many Flemish demonstration projects (Verstraeten *et al.*,  
5 2003). Only when farmers are fully involved in project design and dissemination can new conservation  
6 techniques be successfully implemented on a broader scale. To assist municipalities in establishing a soil  
7 erosion action plan, the Flemish Government subsidises them for organising awareness raising actions such as  
8 erosion-prevention demonstration fields (€125 per field), for excursions with farmers to an erosion prevention  
9 project (€250), for advisory courses (€50 per session) and to approach individual farmers in order to put the  
10 proposed actions into practice (€50 per farmer).

11 Flemish policy makers use several soil erosion indicators (Vandekerckhove *et al.*, 2003). The most  
12 important indicator is the soil erosion risk map of Flanders, based on the application of the RUSLE with  
13 20-m resolution and aggregated at the field scale (Van Rompaey *et al.*, 2000). Furthermore, long-term effects  
14 of soil erosion on soil fertility are evaluated for each field, using soil profile information from the 1:20 000  
15 Belgian Soil Map. Effectiveness of conservation measures is estimated by modelling decreased soil loss at the  
16 field scale and decreased sediment delivery to rivers, using a spatially distributed soil erosion and sediment  
17 delivery model (Verstraeten *et al.*, 2002). The Walloon government is keen to establish a soil erosion control  
18 policy soon, embracing the Flemish model.

### 20 **2.23.2.2 Denmark**

21  
22 Denmark has low erosion rates compared with areas of southern Europe (European Environment Agency,  
23 2000), yet there is a very active programme of soil conservation, particularly in terms of coastal wind erosion  
24 control (Als, 1989). This was initiated during the 19th century, when Hedeselskabet (the Danish Land  
25 Development Service) provided free plants to establish windbreaks to address serious erosion problems.  
26 Regional shelterbelt projects can cover 50 % of expenses for soil preparation, plants, planting and hedge  
27 maintenance for 3 years. Despite the high investment costs, this policy resulted in the planting of 1 000 km of  
28 shelterbelts per year (Veihe *et al.*, 2003). Today, Hedeselskabet continues to offer technical support for  
29 establishing windbreaks, but government subsidies are available for the purchase of certified plants. The  
30 success of these schemes is attributed to farmer participation and good products, plus government involve-  
31 ment, by actioning a specific windbreak law in 1976, which has been revised several times (Veihe *et al.*, 2003).

32 Tillage erosion rates are relatively high in Denmark, at an estimated average of  $\sim 6 \text{ t ha}^{-1}$  (Heckrath, 2000).  
33 The Foulum Research Centre started a 3-year EU-funded project (FAIR3) in 1997, with the aim of quantifying  
34 tillage erosion, determining its effect on soil fertility and investigating the influence of farming practices on  
35 soil erosion and quality (Djurhuus and Heckrath, 2000). Yet there still remains no law dealing specifically with  
36 tillage erosion (Veihe *et al.*, 2003).

37 In 1986, the Danish Nature Council and the national media publicly highlighted eutrophication problems,  
38 which, supported by the Agricultural Council, forced the government to inaugurate the Water Environment  
39 Protection Plan I (WWPP I) and the NPo (nitrogen, phosphorus and organic matter) research programme.  
40 However, policy measures to address water erosion are initiated through the designation of 'Specifically  
41 Vulnerable Agricultural Areas' (SVAA). Government subsidy measures include set-aside, 2-m wide buffer  
42 strips around all watercourses in rural areas and use of rye-grass catch crops (Sibbesen and Iversen, 1997).  
43 WWPP I was superseded in 1998 by the Water Environment Protection Plan II (WWPP II), after erosion  
44 control became recognised as a mechanism for reducing agrochemical contamination.

45 It has been argued that the Danish Land Development Service (Hedeselskabet) could act as a general  
46 organisational model for Europe (Morgan and Rickson, 1990), as they operate an integrated approach to  
47 landscape management (Dubgaard, 2000). However, a change of government in late 2001, associated with

1 major environmental sector cut-backs, has been accompanied by an attempt to simplify laws to ameliorate  
2 working conditions and increase profit within the agricultural sector. This is instigated by the notion that  
3 farmers are best placed to formulate their own decisions regarding farming practices and concomitant  
4 environmental effects (Veihe *et al.*, 2003).

### 6 2.23.2.3 France

7  
8 Many areas of France have a long history of soil erosion problems. For instance, in mountainous areas a state  
9 organisation (Mountain Land Restoration of the National Office of Forests) has been responsible for soil  
10 protection for over a century (Lilin, 1986). Elsewhere, in the Bourgogne vineyards, 19th century private  
11 contracts forced farmers to retrieve and replace eroded soil (Durousset, 1994). In rural areas, no specific  
12 contracts or laws were applied to soil protection and a national survey in 1950 revealed soil erosion in  
13 mountainous areas had been alleviated and insignificant erosion problems existed on farmlands (Hénin and  
14 Gobillot, 1950). This situation prevailed until the 1970s.

15 The first French agricultural law (1960), preceding the first common agricultural policy (1962), led to  
16 conversion of grasslands to cereals and field enlargement (Vivier *et al.*, 1985). Increases in runoff and muddy  
17 floods have also been linked to post-War urbanization. These storm events often had disastrous and sometimes  
18 fatal consequences and caused increased infrastructure damage. A proposed solution was a natural disaster law  
19 (1982), aimed at providing compensation for the victims of natural disasters, but this failed to address the  
20 source of the problem.

21 There continues to be no specific legislation on soil protection in France, because soil erosion is an  
22 unrecognised problem for and by farmers. Moreover, widespread chemical use or irrigation systems are  
23 considered to balance negative impacts of soil erosion on agricultural production. Instead of soil protection,  
24 water (quality and quantity) is viewed as the principal national environmental issue. However, water policies  
25 have some positive effects on erosion control. The Water Law (1992) initiated the division of France according  
26 to the six main catchment areas and established the foundations of the coordinated and sustainable  
27 development of water resources. The Water Law significantly increased the role and responsibilities of  
28 local authorities, reinforced regulations and provided new tools (Schéma Directeur d'Aménagement et de  
29 Gestion des Eaux, SDAGE). In this new framework, the designation of protected areas around catchments  
30 ensured the compulsory reintroduction of permanent grassland and helped local stakeholders manage  
31 problems at the catchment scale instead of at the water resource (i.e. river, lake) level. The framework  
32 considerably assists in dealing with erosion issues. At the same time, the Common Agricultural Policy (CAP)  
33 reforms (European Regulation No. 2078/92) introduced some positive agri-environmental measures for  
34 erosion control, mostly those concerning both protection and extension of grassland. Measures for extensifica-  
35 tion of livestock help to maintain grassland in highland areas. Other long-term (i.e. 20 years or longer)  
36 measures included set-aside of agricultural land (payment €457 ha<sup>-1</sup>) and the reversion of arable land to  
37 extensive grassland (payment €381 ha<sup>-1</sup>). However, in areas of intensive agriculture, these measures were  
38 unattractive. For instance, Upper Normandy gave a supplement of €76 ha<sup>-1</sup> to encourage the reversion of  
39 arable land to extensive grassland for 5 years. In this erosion prevention instrument, grasslands must be located  
40 in thalwegs. Farmers received subsidies corresponding to the long-term set-aside of agricultural land without  
41 being required to commit themselves for 20 years. In 1999, the French farm framework law No. 99-574  
42 introduced 'Territorial Exploitation Contracts' (CTEs) to establish rural development projects that addressed  
43 environmental, economic and social issues. Access to agri-environmental measures became conditional on  
44 agreeing to a CTE. Depending on local needs, some CTE have focused on erosion control, but difficulties in  
45 implementation and high cost led to the halt of CTE measures. In 2001, CTEs were replaced by a new contract  
46 ('Contrat d'Agriculture Durable', CAD), which is more economical since it only considers environmental  
47 issues.

1 Despite the lack of national soil protection legislation, local public authorities became aware that erosion  
2 control involves coordination (Cartier, 2002). Consequently, some authorities formulated plans to solve  
3 environmental problems, notably erosion. For example, recent erosion problems in Seine-Maritime led to the  
4 designation of 'Syndicats de bassin versants' (hereafter referred to as the Syndicates), corresponding to the  
5 main catchment areas. In accordance with the State–Region Plan, Water Authorities control the steering  
6 committee of Syndicates. Syndicates recruit advisors, with hydraulic or agricultural abilities, charged with  
7 implementing hydraulic solutions and encouraging farmers to change agricultural practices.

8 At the local level, it became evident that both hydraulic and agricultural solutions to soil erosion problems  
9 were necessary. However, modifying agricultural practices is difficult, for three reasons. First, farmers'  
10 decisions are strongly influenced by technological and commercial considerations and economic support  
11 under CAP, plus changes in crop management systems have fuelled environmental uncertainties (Souchère  
12 *et al.*, 2003). Second, the diversity of natural conditions (e.g. climate, soils), crops and agricultural practices  
13 causes difficulties in establishing and adapting technical guidelines for agricultural practices suitable for these  
14 different conditions (Papy *et al.*, 1996; Martin *et al.*, 2004). Finally, since agricultural extension services are  
15 directly financed by farmers' taxes to advise farmers how to increase their net income, soil erosion was not  
16 viewed as a farming problem until recently. The main reason is that agricultural advisors experienced  
17 difficulties in estimating the efficiency of new cultivation techniques and also their economic consequences  
18 (Cattan and Mermet, 1998). Furthermore, the incentive for advisors to promote change fails when agricultural  
19 prices decline, because their own funding is reduced. In future, it is hoped that local communities may  
20 subsidise agricultural extension services to contribute to the development of adapted technical guidelines, in  
21 cooperation with the National Agronomy Research Institute (INRA).

#### 22 23 **2.23.2.4 Germany**

24 In Germany, soil conservation legislation has recently been developed, with the main stress on conserving  
25 water quality (Weingarten and Frohberg, 2000). For instance, some soil conservation requirements were  
26 incorporated into the state law of the German Land of Baden-Württemberg in 1991 (Jäger, 1994). Now, over  
27 30 Federal laws and regulations govern environmental protection. The Federal Soil Protection Act (Act on  
28 Protection Against Harmful Changes to Soil and on Rehabilitation of Contaminated Sites) became Federal  
29 Law on 1 March 1999 (Frielinghaus, 2001). The purposes of the Act are to protect and restore soil functions on  
30 a permanent and sustainable basis and to concentrate on both precautions and hazard protection. The latter is  
31 achieved by the provision of a code of good management practice, dealing especially with risks of soil erosion  
32 and compaction. Good farming practice includes designing appropriate treatments with regard to the site and  
33 climatic conditions, maintaining and improving soil structure, promoting biological activity and protecting soil  
34 humus from deterioration. The Act provides a useful model for European soil protection legislation and is  
35 being consulted in the design of legal instruments in other European countries (e.g. Belgium and the UK).

#### 36 37 38 **2.23.2.5 Hungary**

39 Soil erosion is a major soil degradation process in Hungary (Várallyay, 1989; Kertész 2001). Soil protection  
40 legislation includes many forms of soil degradation, such as soil erosion, salinity, extreme acidity and soil  
41 pollution. Before the political regime change of 1989, Hungarian agriculture was highly subsidised, making it  
42 one of the most successful branches of the national economy, with high crop yields and a good reputation,  
43 which was exceptional among the former socialist countries.

44 Soil conservation legislation is enshrined in Chapter VI of Hungarian State Law No. 55 (1994), in which soil  
45 conservation is considered a joint task between the State and the land user. The State develops national soil  
46 conservation policies, ascribes to international agreements and provides farmers and planners with information  
47

1 systems on soil quality, in the form of maps, soil data and legal, economic and technical instruments. A soil  
2 conservation service has been initiated ('Service for Plant Protection and Soil Conservation'), comprised of a  
3 national network of plant health and soil conservation stations, with one in each of the 19 Hungarian counties.

4 Land users are obliged to conserve topsoil and its organic matter content and to conduct environmentally  
5 friendly farm management. This is achieved by appropriate land use, cultivation technology and soil  
6 conservation measures. Law 55 includes actions for eroded and erodible soils, including land-use regulations,  
7 provision of protective vegetal cover and cultivation methods. If these measures cannot guarantee good  
8 protection against erosion, a conservation plan must be formulated, specifying appropriate actions.

9 After the regime change, a complicated reprivatisation process started, resulting in a very complex land  
10 ownership structure. Some cooperative farms ceased to exist, while some remained with changes to their legal  
11 status and many new private farms were established. These were diverse, in terms of both size and financial  
12 viability. Consequently, most new farms are facing challenging economic positions, so soil conservation is  
13 given low priority, which is a recurrent problem in the other former socialist countries.

14 Law 55 stresses that both land users and the State are important partners in soil protection. It is imperative to  
15 learn farmers' opinions on soil erosion and conservation. According to a case study of 44 farmers in Keszthely,  
16 near Lake Balaton, two-thirds of farmers would not spend money on erosion control, even if they had the  
17 resources (Puskás, 2002). However, approximately the same proportion of questioned farmers were convinced  
18 about the damaging effect of erosion and 40 % of them considered soil loss due to erosion was very high.  
19 Some 50 % of the farmers did apply soil conservation measures and only 25 % remained passive and did not  
20 take any conservation action.

21 Joining the EU in 2004 will certainly change both agricultural policy and the attitude of farmers in all  
22 Accession States. The EU strategy on soil conservation will influence the national policies of these countries,  
23 hopefully in a positive direction.

#### 25 **2.23.2.6 Iceland**

26  
27 The severity of soil erosion in Iceland prompted the creation of Northern Europe's only designated soil  
28 conservation service, the Landgrædsla Ríkisins (SCS), in 1907 (Runolfsson, 1978; Arnalds, 2000). The first  
29 50 years were mainly devoted to halting sand dune encroachment and other forms of catastrophic soil erosion  
30 that were threatening several settlements. The Forestry Service, originally established by the same law as the  
31 SCS, had the role of combating woodland destruction and promoting reforestation. It is possible to identify  
32 several attributes of successful practice from the almost 100 years of combating ecosystem degeneration and  
33 desertification in Iceland.

34 About half of the vegetation and 95 % of woodland cover has been lost in the 1100 years of settlement, and  
35 the condition of the remaining vegetation is severely degraded in many areas. Serious soil erosion  
36 characterises 40 % of Iceland according to a national survey completed in 1997 (Arnalds *et al.*, 2001). This  
37 survey represents a major breakthrough, as increasingly, people accept the reality of erosion problems and  
38 focus on their solution.

39 Since 1990, there has been an increasing participatory approach to soil conservation, which has markedly  
40 increased the adoption and success of conservation projects (Arnalds, 1999). The 'Farmer's Reclaim the Land'  
41 Project (Arnalds, 2000) includes a 'cost-share' partnership with farmers, with conservation work jointly  
42 funded by government and farmers. This 'bottom-up' approach encourages involvement and individual  
43 ownership of conservation projects. The rapidly increasing forestry in Iceland, which also has a large role in  
44 land improvements, has a strong farmer and public participation focus.

45 Current developments include a Parliament-approved programme, which gives the SCS an operational  
46 framework for the period 2003–14, with increased funding. The main goals are mitigation of land degradation  
47 and desertification, revegetation of eroded land and attaining sustainable land use. Tools include increased

1 knowledge on problems and solutions, education and advice, land-user responsibility, legal improvements and  
2 widening participation. Carbon sequestration as a tool in meeting Iceland's obligations under the Kyoto  
3 Protocol has also become a major incentive for restoring land health (Arnalds, 2004).

4 Grazing by sheep and horses is a major determinant of land health in Iceland. In the heavily subsidised  
5 sheep industry (with  $\geq 50\%$  of farmers' income being subsidised), the current contract between sheep  
6 producers and government has a cross-compliance clause that, starting in 2003–04, farmers must verify the  
7 ecological sustainability of their operation to the Landgraedsla Ríkisins in order to obtain a full subsidy  
8 (Arnalds and Barkarson, 2003). In the important horse production sector, a voluntary 'bottom-up' quality  
9 control of sustainability is emerging. A new cooperative programme of farm planning, 'Better farms'  
10 combines the forces of soil conservation, forestry, extension and nature conservation in aiding land users in  
11 producing their own property plans. This is a very promising programme, which not only brings farmers into  
12 the planning process, but also improves coordination between the various institutions and organizations  
13 working with farmers. Icelandic soil conservation systems have been presented as an organizational model  
14 generally within Europe (Morgan and Rickson, 1990; Boardman *et al.*, 2003).

### 16 2.23.2.7 Italy

17  
18 Soil erosion plays an important role in Italian land degradation and the European Soil Bureau estimates that  
19  $\sim 70\%$  is at risk of accelerated erosion, defined as erosion  $> 5 \text{ t ha}^{-1} \text{ yr}^{-1}$  (Grimm *et al.*, 2002). These risks  
20 relate to the widespread presence of steep erodible slopes, combined with intensive cultivation and lack of  
21 conservation measures.

22 National legislation considers soil conservation the result of public actions for soil stability and hydraulic  
23 security applied to soil, subsoil, water, urban areas and infrastructures. Italian Law 183 (18 May 1989) on 'soil  
24 defence' identifies a hydrological basin as the operating context for soil conservation planning. Each  
25 hydrological basin is defined by the law as a 'unitary ecosystem', i.e. a complex environment with its own  
26 homogeneity, where coordination and harmonisation of several functions related to soil conservation can be  
27 implemented, along with water management.

28 Establishment of the 'Basin Authorities', at national, regional and interregional levels, is one of the major  
29 innovations brought about by soil conservation reform. They assume responsibility for the territorial  
30 coordination of all the functions carried out by the State, regions and provinces with reference to matters  
31 listed in the law and this is accomplished mainly through 'Basin plans'. These include responsibility for soil  
32 stabilisation by both agrarian and forest hydraulic conservation structures, protection and consolidation of  
33 slopes against landslides and avalanches and prevention of both soil subsidence and seawater intrusion into  
34 aquifers and rivers.

35 Significant progress was made in soil conservation in 1999, when the 'Code of good agricultural policy' was  
36 approved by the Ministry of Agricultural and Forestry Policies. This contains national standards and directions  
37 that can be incorporated by regions and provinces to suit local requirements. The Code was approved under  
38 Article 4 of Directive 91/676/EC of the Council (12 December 1991) and implemented EC Regulation 1257/99.  
39 This regulation calls for agri-environmental measures to finance those actions extending beyond normal 'Good  
40 Agricultural Practice' (GAPn). Regions have defined GAPn to suit different local requirements and have  
41 developed detailed plans to decrease soil erosion risk (particularly decreasing soil tillage and mouldboard  
42 ploughing and promoting crop rotations that ensure soil cover during rainy seasons). As a result, farmers  
43 implementing CAP agri-environmental measures are given subsidies.

44 Other important legislation tools for soil conservation are Law 97 (1994) and the National Programme to  
45 Combat Drought and Desertification (NAP) (1999). Law 97 determines conditions necessary to enhance soil  
46 protection in mountain areas, through the development of farming activities that avoid soil erosion and  
47 maintain the landscape and natural resources. NAP encourages Regions and Basin Authorities to identify

1 'areas vulnerable to desertification' and 'remedial criteria', which includes the prevention and remediation of  
2 soil erosion and salinisation, slope protection through low-impact environmental actions, fire prevention and  
3 the adoption of farming, livestock and forestry management systems aimed at preventing the physical,  
4 chemical and biological degradation of soil.

5 Of the €146.26 million committed to implement NAP, 30 % is devoted to soil protection actions, 15 % to  
6 reducing the impact of industrial activities and 20 % to restoring soil equilibrium. NAP measures will continue  
7 to be adopted through existing funding instruments, such as rural development, water protection and river  
8 basin management plans, forest and afforestation programmes, EU Structural funds (2000–06) and CAP  
9 measures.

### 11 **2.23.2.8 The Netherlands**

12  
13 During recent decades, soil protection legislation has primarily focused on prevention and remediation of soil  
14 contamination. The National Soil Protection Act (1987) provided the framework for establishing rules relating  
15 to categories of activities hazardous to soil. In addition, there are provincial environmental policy plans  
16 devoted to the protection of potable groundwater. Consequently, areas have been designated around extraction  
17 points within which activities hazardous to the soil require particular care. Furthermore, local authority  
18 environmental policy plans are in force and address specific soil protection issues. Additionally, agreements  
19 have been defined and formalised with several branches of industry, including agriculture, to limit the burden  
20 on the environment, and soil in particular. For instance, some soil conservation requirements have been  
21 incorporated by a ruling from the Dutch Agricultural Board (Landbouwschap) in 1992 (Boardman *et al.*,  
22 1994). Soil conservation organisational structures are developing in South Limburg, where a coordinated soil  
23 conservation project was conducted in the 1990s (Duijsings, 1994). The Limburg Soil Conservation Project  
24 (Erosienormeringsprojekt) involves active collaboration between government (provincial and municipality),  
25 agricultural advisory services, university research institutes and local farmers. Demonstration projects and  
26 information dissemination are important components of the programme (Boardman *et al.*, 1994; de Roo *et al.*,  
27 1995). The soil conservation approach followed in the Limburg Soil Conservation Project has been applied  
28 elsewhere in The Netherlands, for instance in defining best conservation strategies for the erosion-prone area  
29 of Groesbeek in the eastern Netherlands (Stolte and Ritsema, 2001; Stolte *et al.*, 2002). Both the Limburg and  
30 Groesbeek Projects may well act as future models for European soil conservation policy.

31 Various national Ministries and related organisations have recently acknowledged that soil legislation and  
32 regulations are too narrowly focused and not harmonised with existing national and international environ-  
33 mental acts. Hence legislation should be reviewed, to make it more transparent and acceptable to the public,  
34 and more easily applicable and manageable for the respective authorities. In 2002, the Ministry of Housing,  
35 Spatial Planning and the Environment (VROM) established a Steering Committee 'Soil' (Stubo), in which  
36 other Ministries, provinces and municipalities participate in preparing a National Integrated Soil Policy  
37 Framework. This new policy should address soil protection in the widest sense, integrating the entire spectrum  
38 of soil physical, chemical and biological functions and properties. Through this process, soil conservation and  
39 runoff and erosion processes may receive appropriate governmental and agency responses.

### 41 **2.23.2.9 Spain**

42  
43 Erosion has long been recognized in Spain (Mallada, 1890), but targeted research did not start until the second  
44 half of the 20th century (see Chapter 1.26), with the creation of the 'Servicio Central de Conservación de  
45 Suelos' (Central Service of Soil Conservation) in 1955. However, since adoption of the 1978 Constitution,  
46 Spain has become less centralized, comprising of 17 Autonomous Regional Governments with many  
47 responsibilities, including erosion control. Many soil erosion and conservation issues are administered by

1 different institutions within the Ministry of the Environment (Dirección General de Conservación de la  
2 Naturaleza, DGCN), formerly ICONA, and the 10 Basin Authorities (Confederaciones Hidrográficas).

3 Autonomous governments and national research institutions [e.g. CSIC, ‘Consejo Superior de Investiga-  
4 ciones Científicas’ (High Research Council for Scientific Research), and CEDEX (‘Centro de Estudios y  
5 Experimentación de Obras Públicas’, Centre for Experimental Studies of Public Works)] and universities have  
6 been the primary contributors to erosion research over the two last decades. In 1981, ICONA, in collaboration  
7 with CSIC and several universities, established the ongoing LUCDEME (Lucha contra la desertificación en el  
8 Mediterráneo) Project, to combat desertification in Mediterranean basins. Since then, maps of actual and  
9 potential soil erosion have been produced. Derived from LUCDEME, in 1995, RESEL [Red de Estaciones de  
10 Evaluación y Seguimiento de la Erosión y la Desertificación en España (Network of Experimental Stations for  
11 Assessing and Monitoring Soil Erosion and Desertification in Spain)] was established. The Project investigates  
12 erosion in problematic environments using field sites, at plot, hillslope and small catchment scales (Rojo and  
13 Sanchez-Fuster, 1997).

14 The National Research and Development Plan is the basic instrument for Spanish research policy and  
15 includes two National Programmes dealing directly with the causes, consequences and control of both  
16 desertification and soil erosion. For the last 20 years, soil erosion research has been conducted and funded by  
17 both national and European institutions. The EC has also given priority to large erosion research projects, such  
18 as DESERTLINKS, MEDACTION, MEDALUS, GEORANGE and PESERA.

19 Spanish agricultural policy, since joining the EC, has been dominated by the CAP, whose priority is the  
20 development of sustainable and ecologically sound agriculture. However, CAP implementation in southern  
21 Mediterranean European countries, and particularly in Spain, has had contradictory effects on soil erosion (e.g.  
22 set-aside resulted in a dramatic increase in the extent of bare soils vulnerable to erosion). CAP reforms, in the  
23 context of Agenda 2000, signified a new era of integrated development of rural areas, including soil protection  
24 and erosion control. Satisfying the new goals of the CAP, Agenda 2000 proposed related changes in the  
25 market-oriented agricultural and rural development policies.

26 A major development of compensatory payments is an option for countries to make direct payments, subject  
27 to compliance with environmental standards (EC1259/99). The Spanish eco-conditionality principle, adopted  
28 by Royal Decree 1322/2002, includes soil conservation practices for protecting fallow land from erosion,  
29 avoidance of stubble burning and prohibition of tillage on steep slopes.

30 Rural Development Policy Regulation 1257/99 provides funding from the European Agricultural  
31 Guidance and Guarantee Fund (EAGGF) to establish new development rural plans for Good Farming  
32 Practices (GFP). The Code of Good Farming Practices was approved by National Decree 4/2001  
33 (13 January 2001). It contains valid national standards for tillage, alternate crops and the use of fertilizers  
34 among other aspects. These standards could be modified by regional governments to suit local require-  
35 ments. Application of agri-environmental measures, regulated by National Decrees (4/2001 and 708/2002),  
36 pursue programmes of environmentally friendly agricultural soil protection practices. However, until  
37 recently, lack of coordination and co-finances meant bureaucratic delays in these programmes. Actual  
38 funded agri-environmental measures (€378.26 million over the period 1995–99) were less than the  
39 estimated mean annual cost of €258.4 million [Ministerio de Agricultura, Pesca y Alimentación  
40 (MAPA), 2000]. Since then, public investment of €1207 million (with an EU contribution of €827 million)  
41 million) has been committed for funding agri-environmental measures for 2000–06. Accompanying CAP  
42 measures are also scheduled and national funding committed.

43 A further instrument of policy development is the afforestation of agricultural lands, regulated by Royal  
44 Decree 6/2001, which substitutes former legislation of Regulation 2079/92, and establishes an Aid Programme  
45 for the afforestation of former agricultural lands. Its main objective is to increase the income of farmers  
46 affected by EU policies, by promoting forestry as an alternative use, to enhance long-term forest resources and  
47 to combat soil erosion and desertification (Gomez-Jover, 1996). In this case, programme implementation is the

1 responsibility of the Autonomous Communities, but is also co-financed by the central administration and the  
2 EU (EAGFF).

3 Since its creation in 1901, the Hydrology and Forest Divisions have been responsible for Spanish forest  
4 administration and have considerable experience in protecting soil against water erosion and rehabilitating  
5 degraded vegetal cover. Since their creation, most responsibility for forest resources and nature conservation  
6 has been transferred to autonomous communities. However, central government [through the General  
7 Directorate for Nature Conservation (DGCN) of the Ministry of the Environment] coordinates plans and  
8 programmes related to soil protection and desertification control. This includes the Spanish Forestry Strategy  
9 (EFE), the National Forest Plan (NFP) and the Autonomous Forest Plan. Objectives include the design of new  
10 forestry policies, emphasising multiple use of forests; organisation of forest resource responsibilities within the  
11 central administration, coordination with autonomous administrations and renewal of forest activity as a means  
12 of generating employment and economic activity.

13 The Forestry Strategy, developed by the NFP, promotes reforestation of degraded plant cover areas, wildfire  
14 protection and application of silvicultural treatments to improve forest quality [Ministerio de Medio Ambiente  
15 (MIMAM), 2001]. The Strategy foresees the reforestation of  $3.8 \times 10^6$  ha affected by soil losses  $>5 \text{ t ha}^{-1} \text{ yr}^{-1}$   
16 and the construction of  $6.9 \times 10^6 \text{ m}^3$  of control structures to stabilize torrential water courses. To address  
17 the application of programmes and actions to control soil erosion and combat desertification the NFP includes  
18 the 'National Plan of Prioritized Actions to Hydrological and Forest Restoration, Soil Erosion Control and  
19 Combating Desertification'.

20 Thirteen Autonomous Communities have drafted documents on forest policy and forestry master plans, of  
21 which seven have promulgated regional forest resource laws. Most plans identify the restoration of degraded  
22 lands affected by soil erosion, by the means of revegetation, and the need for sustainable forestry to control  
23 soil erosion and its off-site effects. For instance, the Forest Law of Valencia (Ley Forestal de la Comunidad  
24 Valenciana 3/93) highlights the fight against desertification, prevention of forest fires, restoration of vegetable  
25 cover and flood protection as major objectives. Thus, a reforestation plan was implemented for 1994–99. The  
26 plan foresaw re-vegetation of 100,000 ha of public lands, with a budget of €160 million.

27 The 'National Plan of Prioritized Actions to Hydrological and Forest Restoration, Soil Erosion Control and  
28 Combating Desertification' promoted by DGCN in 1991 (Rojo, 1998) aims to control soil erosion and  
29 establish locations and priorities for land rehabilitation in degraded river basins, according to both the use of  
30 river basins as operational units and soil erosion severity. Hydrological and forest restoration activities  
31 developed under 'Cooperation Conventions with Autonomous Government' (1997–2001) shows that the Plan  
32 is failing its initial goals.

33 In the future, 'The National Plan of Hydrological Forest Restoration and Erosion Control' will be  
34 integrated in the National Actions Programme to Combat Desertification (NAPD). When Spain became a  
35 signatory to the UN Convention to Combat Desertification (CCD) in 1996, it was committed to establishing  
36 an NAPD. This is being prepared by a working party group coordinated by DGCCANA, with the main  
37 objective of sustainable development in Spain's arid, semi-arid and dry subhumid areas and, particularly,  
38 the prevention and reduction of land degradation, the rehabilitation of partly degraded land and reclamation  
39 of desertified areas.

40 In 2002, DGCN initiated a national inventory of soil erosion (INES), which aims to locate, quantify and  
41 analyse the evolution of erosion processes in Spain. The main objectives are to prioritise target areas for soil  
42 conservation and define and evaluate national plans on reforestation, plant cover improvement and the  
43 management of forest biodiversity. In every province an inventory and map (1:50 000 scale) will be made of  
44 rill, gully, river bank and wind erosion and mass movements. Inventory maps have been completed for the  
45 provinces of Madrid, Murcia and Lugo, five more are at an advanced stage and another five are in progress.  
46 The inventory will cover all 50 provinces and will take about 10 years to complete. The inventory will then be  
47 repeated on a cyclical basis.

### 2.23.2.10 The UK

There is increasing evidence in the UK that soil erosion poses at least a moderate to severe problem at local to regional scales (Boardman and Evans, 1994), although government involvement remains only moderate. In the 1980s, the Agricultural Development and Advisory Service (ADAS), then a constituent part of the Ministry of Agriculture, Fisheries and Food (MAFF), published two advisory bulletins to help farmers minimize erosion (ADAS, 1984, 1985). Since then, a 'Code of Good Agricultural Practice for the Protection of Soil' has also been published (MAFF, 1993) and a detailed guide to soil conservation for farmers (MAFF, 1999). The Environment Agency (formerly the National Rivers Authority) was established on 1 April 1996, with a remit to protect the UK environment, including its soils. It is likely the Environment Agency will play a significant role in soil conservation (Boardman, 2002).

Several relatively simple approaches would reduce national soil erosion. MAFF recognised that set-aside schemes may contribute to soil conservation (MAFF, 1991, 1994, 1999; Margach, 1993) and, furthermore, grassland protection of riparian zones would decrease sediment delivery to streams and rivers (Morgan, 1992). The MAFF (1999) booklets advised farmers on simple techniques to produce a farm-size soil erosion risk map and approaches to combat erosion on areas 'at risk'. Particular emphasis was placed on grass strips on arable slopes and buffer strips in riparian zones. Grass strips were recommended to be 5–15 m wide and every 50–150 m downslope, with the width of the strip increasing and the gap between strips decreasing as erosion risk increased.

An independent, government-sponsored report advised the government to devise and implement a soil protection policy (Royal Commission, 1996). An outcome of the report was the preparation of a 'Draft Soil Strategy for England – a Consultation Paper' in March 2001 [Department of the Environment, Transport and the Regions (DETR), 2001]. Following public consultation, in which interested members of the public were invited to comment on the draft, the Strategy is being prepared for approval by the UK Parliament. By comparison, the nature and scope of the Strategy is similar to the German 'Federal Soil Protection Act (1999)'. In the more federal government structure, recently adopted in the UK, the legislation will be specifically for England, although legislation will be forwarded to the separate regional assemblies of Scotland, Wales and Northern Ireland. Although this will facilitate legislation fine-tuned to local conditions, it will inevitably slow the process of implementation, as fundamentally similar legislation will be subject to ratification by four separate legislative assemblies. The draft strategy is available at: <http://www.defra.gov.uk/environment/consult/dss/>.

Progress has been made, with the submission of two reports. Stirling University (Scotland) reported to the Scottish Executive, recommending a soil protection strategy for Scotland (Adderley *et al.*, 2001). The Centre for Ecology and Hydrology (Bangor, Wales) has presented a similar strategy to the Welsh Assembly. The respective texts are available at <http://www.envsci.stir.ac.uk/spstrategy/index.htm> and <http://www.bangor.ceh.ac.uk/English/science/reports.htm>.

Agri-environment schemes in the UK, directed by MAFF [renamed the Department for Environment, Food and Rural Affairs (DEFRA) in 2001], aim to secure environmental benefits above those of Good Farming Practice and cross-compliance conditions. Introduced in 1987 to implement EU Council Regulation 797/85, they were designed to prevent loss of habitat and landscape features associated with intensification at sites targeted by the Environmentally Sensitive Areas (ESA) Scheme. Subsequently, in 1991, the Countryside Steward Scheme (CSS) was established to provide incentives to landowners, farmers and other land managers to take specific measures to conserve, enhance and/or recreate important landscape types. In 1994, the Habitat Scheme (HS) was initiated to create, protect and enhance wildlife habitats by removing land from agricultural production and promoting environmentally sound land management practices. In 1995, the Moorland Scheme (MS) was launched with the objective of protecting and improving the upland moorland environment. In 1998, the Arable Stewardship Pilot

1 Scheme (ASPS) was created to assess alternative arable management options for conserving and enhancing  
2 farmland biodiversity (Ecoscope Applied Ecologists, 2003). In December 2003, the UK government  
3 initiated a new agri-environment initiative, known as the 'Environmental Stewardship Scheme' (ESS). The  
4 scheme operates two-tier participation. The 'Lower Entry Level Scheme' encourages farmers to deliver  
5 simple, yet effective, environmental land management, which will help to diffuse pollution, prevent loss of  
6 biodiversity and landscape character and damage to the historic environment. The 'Higher Level Scheme',  
7 launched in 2005, is based on the existing CSS and ESAs. Its five main objectives are wildlife conservation;  
8 protection of the historic environment; maintenance and enhancement of landscape quality and character;  
9 promoting public access; and understanding and resource protection ([http://www.defra.gov.uk/erdp/  
10 reviews/agrienv/default.htm](http://www.defra.gov.uk/erdp/reviews/agrienv/default.htm)).

11 DEFRA is drafting 'The First Soil Action Plan for England 2004–06'. The plan has the overarching  
12 principle to 'ensure that England's soils will be protected and managed to optimise the varied functions that  
13 soils perform for society, in keeping with the principles of sustainable development and on the basis of sound  
14 evidence' (draft 7 November 2003).

## 17 **2.23.3 REVIEW OF RELEVANT NATIONAL POLICIES BEYOND EUROPE**

### 19 **2.23.3.1 Australia**

21 The Australian Landcare system offers a possible model for group participation in soil conservation.  
22 Landcare began in Victoria in 1986 and has grown to encompass 25 % of the farming community  
23 (Campbell, 1995). Landcare adopts an integrated and holistic approach to resource sustainability and is a  
24 cooperative venture between Federal and State governments, extension services, consultants and farmers  
25 (Curtis and DeLacy, 1995). Over 2700 Landcare groups cooperate on local land degradation issues, which  
26 are usually managed at the catchment scale (Hannam, 2000). Issues addressed include identification of land  
27 degradation problems and implementation of solutions, development of demonstration sites and promotion  
28 of community and stewardship (Ewing, 2000). Education is especially important, in particular, encourage-  
29 ment of 'land literacy' among participants (Campbell, 1995). The scheme has also been successfully  
30 adopted in New Zealand (Ministry for the Environment of New Zealand, 1996; Bettjeman, 2000). Some  
31 have added cautionary notes over the adoption of Landcare schemes, including the need to maintain  
32 fundamental research capabilities (Hannam, 2000) and avoidance of over-optimism on the potential  
33 success of the scheme (Bradsen, 2000).

### 35 **2.23.3.2 North America**

37 In North America, soil erosion research and conservation has been established since the 1930s. A great deal of  
38 experience has been accumulated and, consequently, many North American policies, lessons and perspectives  
39 have direct relevance to the formulation of European soil conservation policies (Boardman, 1991). In the  
40 European context, it should be noted, US and Canadian systems are well structured to deal with erosion  
41 policies at national, regional and local levels. Policies are in place, which are largely advisory, but incorporate  
42 elements of coercion.

43 North American soil conservation was galvanised by the severe erosion of the 'Dust Bowl' of the US Great  
44 Plains States during the 'dirty thirties' (Bagley, 1979; Hurt 1981). The US Soil Erosion Service, founded in  
45 1933, renamed the Soil Conservation Service (SCS) in 1935, is a permanent branch of the US Department of  
46 Agriculture (USDA) (Soil Conservation Society of America, 1980). The US SCS [renamed the National  
47 Resources Conservation Service (NRCS) in 1994] remains the principal authority responsible for soil

1 conservation, with over 13 000 employees, and is directly under the jurisdiction of the Secretary for  
2 Agriculture. The Soil Survey Division of NRCS provides local and regional soil maps, essential for the  
3 formulation of soil conservation strategies. At the NRCS National Soil Survey Laboratory in Lincoln, NE, the  
4 Soil Survey Division also analyses samples collected by the staff of around 1000 scientists, thus ensuring  
5 accurate and comparable data.

6 To administer NRCS policies, soil conservation districts are established, usually adopting County  
7 boundaries (the administrative subunit of the State). Each district has a conservationist, who is responsible  
8 for providing local soil conservation advice. Every district also possesses a 'Board of Supervisors', consisting  
9 of unpaid citizens who share an interest in soil conservation. There are 2950 conservation districts, covering  
10 most of the  $171 \times 10^6$  ha of US cropland (Steiner, 1990). Each State appoints a State conservationist,  
11 responsible to the NRCS headquarters in Washington, DC.

12 The NRCS adopts a voluntary land-use planning approach to soil conservation, termed the 'information,  
13 education and subsidy' approach (Napier and Napier, 2000a). Upon request by farmers, a conservationist will  
14 recommend appropriate strategies for reducing soil erosion to acceptable levels. The NRCS does not possess  
15 mandatory powers of land control, but non-implementation of conservation policies can exclude farmers from  
16 Federal and State grant-aided programmes. The Conservation Title of the Food Security Act (1985) and the  
17 Food, Agriculture, Conservation and Trade Act (1990) introduced elements of coercion into soil conservation  
18 policies (Napier, 1990; Esseks and Kraft, 1991). 'Sodbuster provisions' deny farmers access to farm  
19 programme benefits if they crop 'highly erodible land' (HEL) without an approved soil conservation plan.  
20 'Swampbuster provisions' aim to prevent conversion of wetland to crop production, while the 'Conservation  
21 Reserve Program' (CRP) attempts to retire highly erodible land from agricultural production. The 'Conserva-  
22 tion Compliance Provisions' have the greatest potential for long-term reductions in erosion. According to  
23 legislation, operators of HEL had to develop an officially approved plan by 1 January 1990, with full  
24 implementation by 1 January 1995. Non-compliance resulted in the loss of all USDA farm benefits until an  
25 approved plan was implemented (Napier, 1990). The Federal Agricultural Improvement and Reform Act  
26 (1996), or FAIRA, increased the emphasis on conservation in current food security legislation (Napier and  
27 Napier, 2000a; Weber and Margheim, 2000).

28 Much can be learnt from North American strategies and policies, but they would need modifying and  
29 adapting to European conditions. For instance, set-aside policies were borrowed from North America and  
30 established to reduce European grain surpluses (Marsh, 1987; Clarke, 1992; Jones, 1992). However, it is  
31 feasible that steep to moderate slopes with erodible soils and other vulnerable parts of fields (i.e. depressions,  
32 minor dry valleys and land adjacent to water courses) be put to set-aside (Boardman, 1988; Fullen, 1991a;  
33 Chambers and Garwood, 2000). This policy was successfully adopted in the 'Permanent Cover Program' in the  
34 Canadian Prairies (Vaisey *et al.*, 2000). Preliminary evidence from both the South Downs of England  
35 (R Eevans and J Boardman, personal communication, 1996) and Shropshire, England (Fullen, 1998) suggests  
36 temporary set-aside and permanent grassland are very effective for soil and water conservation. A 'more  
37 directed policy' has been proposed in the UK by the Royal Commission on Environmental Protection, which  
38 recommended 'the government make maximum use of national discretion to adopt environmental and soil  
39 protection criteria in the selection of land for set-aside, and encourage this approach to set-aside at EU level'  
40 (Royal Commission, 1996; Recommendation 24).

41 The impact of set-aside on soil erosion and its off-site consequences is not straightforward, as demonstrated  
42 by a modelling exercise for cultivated catchments in central Belgium (Verstraeten *et al.*, 2002). For regions  
43 with different soil, land-use and morphological characteristics, the impact of the same set-aside percentage  
44 results in different erosion rates. Furthermore, decreases in off-site sediment delivery may be different from  
45 decreases in soil loss, depending on the location and environment of set-aside fields. Sites prone to soil erosion  
46 are not necessarily those fields delivering most sediment to river channels. The optimal location of set-aside  
47 fields will therefore depend on policy goals.

1 In a 50-year review of US soil conservation experience, Sanders (1990) argued that policies must be  
2 developed, based on thorough analysis and understanding of the problems, and they should remedy causes of  
3 erosion, rather than simply treating symptoms.  
4

#### 5 6 **2.23.4 SYNTHESIS OF NATIONAL POLICIES: TOWARDS A SOIL** 7 **CONSERVATION POLICY FOR EUROPE** 8

9 Successes and failures of various policies developed in the industrialised world (Australia, Europe and North  
10 America) were debated at an international conference in Prague in 1996. Based on these discussions, Napier  
11 *et al.* (2000a) presented 38 chapters, which thoroughly explore these issues.

12 Voluntary and non-government organisations have an important role in developing soil conservation  
13 policy (Schnepf, 2000). Development of informed debate is critical to future development of European soil  
14 conservation and several organisational developments are assisting. The European Society for Soil  
15 Conservation (ESSC) was founded in November 1989, with the mission of developing an integrated  
16 European approach to issues of soil erosion and conservation. The ESSC consists of a group of scientists  
17 attempting to influence governmental policies and public attitudes towards erosion problems ([http://](http://www.zalf.de/essc/essc.htm)  
18 [www.zalf.de/essc/essc.htm](http://www.zalf.de/essc/essc.htm)). It consists of 534 members from 46 countries, including 36 European  
19 countries (Valencia Congress, March 2000). The initial ESSC Conference, in Coventry in 1990, resulted  
20 in a valuable overview of European research (Boardman *et al.*, 1990). The first ESSC Congress, held at  
21 Silsoe College, Bedfordshire, UK, resulted in a further review (Rickson, 1994). The second ESSC  
22 Congress, held at Munich, in September 1996, provided selected papers in a special issue of *Soil and*  
23 *Tillage Research* (1998, Volume 46, Issues 1 and 2) (Auerswald and Kutilek, 1998). The third ESSC  
24 Congress, held in Valencia, Spain, in March 2000, focused on highlighting ESSC work for public attention.  
25 It offered a timely review of the status of soil erosion and conservation research at the dawn of the new  
26 millennium, with the publication of two comprehensive Congress volumes (including 196 chapters),  
27 mainly dealing with the European dimension (Rubio *et al.*, 2002). The fourth ESSC Congress was held in  
28 Budapest in May 2004.

29 An e-mail based 'soil erosion discussion group', coordinated by Purdue University, West Lafayette, IN,  
30 USA (e-mail address: [se-list@ecn.purdue.edu](mailto:se-list@ecn.purdue.edu)), provides a valuable means of discussion and communication,  
31 linking over 360 subscribers world-wide (Bernsdorf and Favis-Mortlock, 1995).

32 Increased activity of the International Erosion Control Association (IECA) in Europe is another welcome  
33 trend. The IECA was founded in 1972 and based at Steamboat Springs, CO, USA (<http://www.ieca.org/>). It  
34 operates mainly in the USA, but its first European Conference was held in Barcelona in May 1996. The  
35 practical outlook of the IECA, with strong emphasis on technical, engineering and industrial solutions to  
36 erosion and sediment control, should complement and enhance the activities of more academically orientated  
37 European researchers.

38 Rational policies must be designed at international, national, regional and local scales. At the European  
39 scale, the European Soil Bureau (ESB) is a useful forum for discussion of Europe's soil management  
40 problems. This was initiated by meetings of the heads of Europe's soil survey organisations in 1989 and led to  
41 the establishment of the ESB in 1996, based at the EU Joint Research Centre in Ispra, Italy. The ESB focuses  
42 on harmonising soil survey operations, but could play an increasingly important role in promoting sustainable  
43 development of Europe's soil resources (<http://esb.aris.sai.jrc.it/>).

44 The EC is playing an increasingly active role in developing soil conservation strategies. The 6th  
45 Environmental Action Programme (EAP) of the European Union was adopted in 2002 and identified soil as  
46 a natural system and a non-renewable resource and established the general objective to protect soil,  
47 particularly against erosion and pollution.

1 The Environment Directorate General of the EC has drafted a policy that ‘contains the commitment by the  
2 Commission to develop a thematic strategy for soil with the ultimate goal of raising the political importance of  
3 soil issues at EU level, so soil protection receives adequate attention’. The ‘Soil Protection Communication’  
4 (26 October 2001) passed public consultation enabling the policy statement ‘Towards a Thematic Strategy for  
5 Soil Protection’ to be published on 16 April 2002 (Commission of the European Communities, 2002).

6 The EC is developing a broad overview of soil problems in Europe. It considers erosion to be a critical  
7 problem for Europe’s soils and believes that soil conservation is a cornerstone of soil protection policy. The  
8 Commission recognises that problems in southern Europe are well documented and are clearly evident in the  
9 presence of rills and gullies. However, the situation is more insidious in northern Europe. For instance, high  
10 erosion rates can occur by sheet erosion, which is usually not considered a significant problem by farmers.  
11 Furthermore, associated pollution of waterways by erosion-derived agrochemicals is largely ‘invisible’ and  
12 does not receive sufficient attention. The Commission takes the view that considerable data exist on the nature,  
13 extent and severity of erosion on European agricultural soils. The need for more information and monitoring of  
14 soil erosion processes, such as gully erosion, piping and soil losses associated with root and tuber harvesting  
15 and land levelling, have been identified. The need both to consider the extent and severity of soil degradation  
16 problems and to support attempts to develop suitable and innovative soil conservation measures is stressed,  
17 including the potential of conservation tillage as a soil conservation measure. There is significant concern  
18 about possible implications of the former ‘Accession States’ joining the EU, as many have considerable soil  
19 management problems, particularly associated with chemical contamination (Commission of the European  
20 Communities, 2002).

21 The EC believes that education is important to increase public awareness of soil as a resource and proposes  
22 a glossary of soil terms would be useful. The importance of soil as a store for Europe’s archaeological heritage  
23 is also acknowledged.

24 The EC identifies areas where there is a paucity of information on Europe’s soils. Relatively little is  
25 known about long-term trends in soil organic matter contents, the carbon sequestration potential of soil and  
26 the capacity of soils to safely absorb sludges. Better understanding is required of linkages between erosion-  
27 derived pollution of waterways and threats to the integrity of aquatic ecosystems. There is debate on the  
28 potential effects of ecological–organic agriculture on soils, but it is generally agreed it will increase soil  
29 biodiversity. The Commission appeals for more data on soil sealing and sterilisation by construction and  
30 industrial activities. The considerable heterogeneity of Europe’s soils poses problems in developing  
31 universally applicable policies. Further challenges are posed by European political diversity, with 156  
32 political units (e.g. Lande, States, Cantons) responsible for policy implementation within the former 15 EU  
33 states.

34 It is recognised that although there are some 41 international conventions dealing with the environment,  
35 only one deals with land and is specific to arid environments (UN Convention to Combat Desertification,  
36 UNCCD). The Commission is investigating national policies on soil protection (for instance, in Belgium,  
37 Denmark, France, Germany, The Netherlands, Sweden and the UK) and attempting to abstract different  
38 policy options applicable to Europe. The ‘Thematic Strategy for Soil Protection’ will be updated regularly,  
39 in consultation with interested parties, including the ‘Stakeholders Forum’ established in January 2002.  
40 This strategy is a permanent and integral component of the Commission’s overall environmental and  
41 agricultural policy. The text is available at the EC web site: [http://europa.eu.int/comm/environment/soil/  
42 index.htm](http://europa.eu.int/comm/environment/soil/index.htm).

43 Interested members of the public can express their views on soil policy to the EC via the web-based  
44 ‘CIRCA-Soil’ (the soil policy electronic forum), launched in July 2003, at [http://forum.europa.eu.int/Public/  
45 irc/env/Home/main](http://forum.europa.eu.int/Public/irc/env/Home/main).

46 In spring 2003, the EC established five technical working groups with the main goals of gathering existing  
47 information on the eight identified soil threats and making recommendations for measures and policy options.

1 Three soil priority areas were defined (erosion, contamination and organic matter), constituting the core issues  
2 of three thematic working groups. Further, two horizontal working groups on monitoring and research were  
3 established. The other five soil threats were dealt with by the most related working groups: compaction, floods,  
4 landslides and salinisation were linked to erosion, biodiversity was linked to organic matter and sealing by  
5 construction was linked to research. The final aim of the working groups is to contribute to EC deliverables.  
6 However, what will be delivered depends on political decisions taken by the new Commissioner, appointed in  
7 November 2004. It will be a single package, including a communication on actions, and possibly a Soil  
8 Framework Directive. The EC will also formulate directions for future research.

9 Generally, especially from within the Working Group on Soil Erosion, the development and implemen-  
10 tation of a Soil Framework Directive was supported. This was in order to reach a more coherent and visible  
11 soil policy, rather than only linking up soil issues with different other policies. This would imply legally  
12 binding measures to protect soils from threats and the obligation for Member States to monitor the state and  
13 trends of soils with respect to these threats. The foundation of national and/or a European Soil  
14 Conservation Service was considered necessary for the proper implementation of such a Directive. The  
15 amendment of existing policies to address soil threats better, e.g. through the Water Framework Directive  
16 or the CAP, was also suggested. After the technical working groups finished their reports (15 May 2004),  
17 the consultation phase was finished. The future development of the EU Soil policy will then largely be  
18 determined at the political level, but at the administrative and expert level the protection of soils is  
19 acknowledged as a high-priority concern.

20 Governments have crucial roles to play, especially at a national level (Napier *et al.*, 2000b). They should  
21 direct their policy to monitor the status of their national soils and to achieve proper soil use and conservation.  
22 A commendable example of a clearly formulated policy statement is the 'Sustainable Land Management  
23 Strategy' of the New Zealand Government (Ministry for the Environment of New Zealand, 1996; Bettjeman,  
24 2000). There is debate whether government actions should be on a cooperative and voluntary basis (e.g.  
25 Johnson, 2000), or whether there should be a more coercive and regulatory approach (e.g. Napier and Napier,  
26 2000a,b). Where possible, government involvement should not be authoritarian or punitive, but should aim at  
27 facilitating conservation by assisting in the identification of problems, in tackling the underlying causes of soil  
28 misuse and by encouraging necessary actions. A senior administrative body or commission would be  
29 necessary for such a task. The body should, in consultation with interested parties, establish, promote and  
30 finance research priorities. However, if all reasonable attempts at cooperation fail, then a more coercive  
31 approach would be necessary. Current emphasis on the 'polluter pays' principle goes some way towards  
32 addressing this issue. The responsible body should have clear and verifiable aims. For instance, the 'Ontario  
33 Land Stewardship Program' (OLSP) in Canada has been criticized for lacking specific and measurable aims  
34 (Stonehouse, 2000).

35 Availability of accurate, high-quality soil data is pivotal to a successful policy. A European and national  
36 inventory of land resources is necessary, so gaps in knowledge are identified and, where necessary, studies  
37 commissioned. National soil survey organisations must play a vital role in providing information (Young,  
38 1991). Bullock and Thompson (1996) argued for a two-stage integrated policy for improving the sustainability  
39 of UK soil resources. First is the identification of the current state of soils, to assess the capability,  
40 vulnerability, sensitivity and resilience of soils, to inform the decision-making process. Second is to match  
41 soils and their use, so land use is sustainable and appropriate.

42 European governments have tended to regard national soil survey organisations as rather esoteric  
43 entities, and their finances are often vulnerable to the whims of finance ministries. Much can be learnt from  
44 the US experience, where the Soil Survey is a respected, properly funded organisation, with a relatively  
45 high profile in public awareness (Batie, 1985). European Soil Surveys should consider adopting the policy  
46 of US County soil surveys of incorporating USLE soil erodibility ( $K$ ) value assessments into their mapping  
47 at the series scale. This approach has proved useful for regional erodibility mapping in Belgium (Pauwels

1 *et al.*, 1980), Denmark (Madsen *et al.*, 1986) and Germany (Becher *et al.*, 1980; Jäger, 1994). In addition,  
2 temporal trends in organic content of arable soils should be closely monitored (Fullen, 1991b). The NSRI  
3 evaluated changes in soil organic matter between the early 1980s and 1995 and found that concentrations  
4 are decreasing on arable topsoils and many former peats are now classified as humose mineral soils  
5 (Bullock and Burton, 1996).

6 Paucity of information on the costs of erosion impedes full evaluation of its effects. Data collection and  
7 evaluation of both 'on-site' and 'off-site' costs are problematic. Many costs are difficult to quantify, borne by  
8 various groups (e.g. local councils, water authorities, insurance companies and householders), are inherently  
9 difficult to collate and are not necessarily costed as being directly due to soil erosion (Stammers and  
10 Boardman, 1984). Costing of erosion and related flooding episodes on the South Downs of England placed  
11 'off-site' costs between €90 000 and €350 000 and up to almost €1.8 million at Rottingdean in 1987  
12 (Boardman *et al.*, 1994). A tentative costing of both 'on-site' and 'off-site' erosion in England and Wales  
13 produced a combined total of €1,400 million per year (Evans, 1995). The 'off-site' cost associated with  
14 downstream sedimentation is considered to be the dominant problem (ADAS, 2002; Boardman, 2002). Similar  
15 'off-site', high-cost scenarios of erosion episodes have been suggested in central Belgium, including non-  
16 quantifiable social costs of stress, induced by flood damage and risk to property (Verstraeten and Poesen,  
17 1999). Comprehensive and accurate costing of soil erosion would be helpful, both in evaluating the problem  
18 and in planning policy responses. Adoption of the 'polluter pays' principle would promote more effective  
19 conservation.

20 On a regional scale, skilled personnel are necessary for consultative duties. With reference to the UK,  
21 Boardman (1988, 1991) suggested the establishment of a small soil conservation unit within MAFF.  
22 Morgan and Rickson (1990) argued the Danish Land Development Service (Hedeselskabet) or the  
23 Landgraedsla Rikisins of Iceland could act as organisational models generally within Europe. Establish-  
24 ment of European soil conservation services, whether as distinct entities or as subdivisions of agricultural  
25 advisory services, merits discussion. Essential components of any soil conservation service should supply  
26 free information and advice to agriculturalists and interested bodies. Advisory services should also freely  
27 disseminate information to the public, particularly educational establishments. Current UK government  
28 policy of charging fees for advice from DEFRA is counterproductive. It is imperative soil that conservation  
29 field demonstrations are organised, so farmers can see tangible evidence of conservation benefits. As  
30 erosion occurs on a field scale, local conservation policies are essential (Evans, 1990). Agriculturalists  
31 must be able to call freely upon the advice and expertise of soil scientists. In this respect, Europeans have  
32 much to learn from the US NRCS. Advisors should assist farmers, identifying causes of erosion and  
33 selecting appropriate technologies for remediation. Evolution of a conservation plan should be an  
34 interactive process between an advisor and agriculturalist, leading to the development of a range of  
35 possible costed strategies. Costs of remedial or preventative measures may be prohibitive, therefore the US  
36 cost-share system, where the NRCS can meet up to 75 % of the cost of conservation measures, seems  
37 appropriate. A similar approach is adopted by the Icelandic SCS (Arnalds, 1999), the Canadian OLSP  
38 (Stonehouse, 2000) and by the Ministry of Flanders, Belgium (Vandekerckhove *et al.*, 2003). With respect  
39 to US agriculture, it was argued that 'environmental credits' could be introduced to recognise and  
40 financially reward actions of producers who voluntarily introduce conservation systems on their land  
41 (Weber and Margheim, 2000). Sanders (1990) stressed the need for local voluntary organisations to discuss  
42 erosion problems. The US County soil conservation district could provide a useful model. In such a forum,  
43 interested parties meet and discuss local erosion problems and potential solutions.

44 Education plays an essential role in informing the public on the importance of soil as a resource. Several  
45 illustrative examples of good practice can be identified. These include the US NRCS policy to identify 'State  
46 Soils', that is, a specific soil type associated with each State. These are used in school education programmes,  
47 with students visiting representative profiles. Soil education was a major component of the EXPO 2000

1 Exhibition, near Hannover, Germany, held from June to October 2000 (Auerswald, 2000; <http://www.obe2000.de>). In Devon, England, the Institute of Grassland and Environmental Research (IGER) has made  
2 important contributions, including the production of an educational resource leaflet for teachers and students  
3 entitled 'Working with Soil'. IGER has also developed educational 'Soil Trails', designed on the basis of  
4 'Nature Trails'.  
5

6 It is imperative that the broader benefits for soil conservation to society are recognised. Not only is  
7 conservation beneficial to agricultural development, it also assists several environmental objectives. For  
8 instance, soil conservation is compatible with habitat creation and the promotion of biodiversity. Soil  
9 conservation can assist with the accumulation of soil organic matter, which is an important 'sink' for  
10 atmospheric carbon. In turn, carbon sequestration can help ameliorate global warming (Wedin and Tilman,  
11 1996; Arnalds, 1999; Lal, 2000).  
12

### 13 **2.23.5 CONCLUSIONS**

14 The extent and severity of erosion on European soils have increased markedly over the last 50 years,  
15 particularly on arable land. However, government action and advice on soil conservation have been limited.  
16 Recently, government and agency interest in soil erosion has increased, largely owing to 'off-site' costs and  
17 problems associated with downstream sedimentation and water quality. Taking countries representative of  
18 different regions of Europe, a broad range of instruments are used, usually at the catchment scale. These include  
19 agri-environment measures, legal instruments, erosion control plans, the promotion of participatory approaches,  
20 subsidies, education programmes and the development of advisory services. These instruments are promoted to  
21 varying degrees, in response to both the severity of erosion problems and the degree of government interest and  
22 involvement. In this review of strategies to improve soil conservation, several 'best management' practices are  
23 identified. Policies at international, national, regional and local scales should include:  
24  
25

- 26 1. Initiation of national soil conservation services. These organisations should be properly-funded and  
27 relatively well publicised.
- 28 2. Full mapping, monitoring and costing of erosion risk by national soil survey organisations.
- 29 3. A participatory approach to soil conservation, involving farmers and interested members of the public.
- 30 4. A 'cost-share' partnership between government and farmers.
- 31 5. Development of rational land-use policies, such as targeting temporary and permanent set-aside on steep  
32 and erodible land, use of grass strips on erodible arable slopes and the protection and management of  
33 riparian zones.
- 34 6. Increased public understanding and awareness of the value of soil resources through education pro-  
35 grammes. Education schemes should particularly encourage 'land literacy' among participants.
- 36 7. Broader benefits of effective soil conservation to society should be recognised and promoted, such as its  
37 potential contribution to habitat creation, promotion of biodiversity and carbon sequestration.  
38  
39  
40

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UNCORRECTED PROOFS